

Bioremediation of uranium-contaminated groundwater: a systems approach to subsurface biogeochemistry

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Adding organic electron donors to stimulate microbial reduction of highly soluble U(VI) to less soluble U(IV) is a promising strategy for immobilizing uranium in contaminated subsurface environments. Studies suggest that diagnosing the *in situ* physiological status of the subsurface community during uranium bioremediation with environmental transcriptomic and proteomic techniques can identify factors potentially limiting U(VI) reduction activity. Models which couple genome-scale *in silico* representations of the metabolism of key microbial populations with geochemical and hydrological models may be able to predict the outcome of bioremediation strategies and aid in the development of new approaches. Concerns remain about the long-term stability of sequestered U(IV) minerals and the release of co-contaminants associated with Fe(III) oxides, which might be overcome through targeted delivery of electrons to select microorganisms using *in situ* electrodes.

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Introduction

Uranium contamination of groundwater is extensive worldwide and prohibitively expensive to remediate with traditional strategies, such as pump-and-treat. A promising alternative is to promote microbial reduction of U(VI) to U(IV), which removes uranium from groundwater as an insoluble precipitate (Figure 1) [1[•],2[•],3]. This constitutes a more attractive option than other biological alternatives, such as biosorption or bioaccumulation [4], due to the greater stability of the reduced form [5]. The effec-

tiveness of microbial reduction for removing uranium from contaminated groundwater has been repeatedly validated across a range of environments, including saturated alluvial sediments [2[•]] and fractured saprolite [6], with remediation of environments impacted by uranium mining via *in situ* recovery (ISR) gaining considerable interest [7].

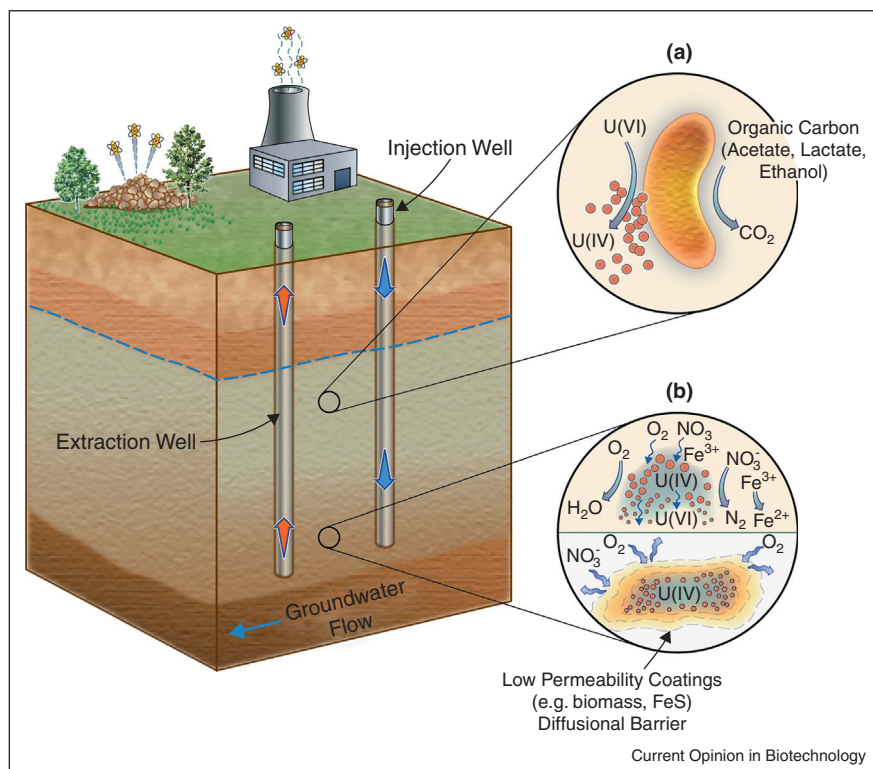
In the twenty years since the discovery of microbial U(VI) reduction, there has been significant progress in understanding the interplay of biology, hydrology, and geochemistry, to the point where in some instances it is possible to predictively model the coupled processes in subsurface environments [8,9]. However, significant questions remain about how to best optimize *in situ* uranium bioremediation because of important knowledge gaps about microbe–microbe and microbe–mineral interactions, as well as a need for more information regarding uranium geochemistry in subsurface environments.

Microorganisms associated with subsurface U(VI) reduction

A broad diversity of microorganisms available in pure culture is capable of U(VI) reduction (Figure 2) [1[•],4]. Most are anaerobes that have the ability to reduce other metals, most notably Fe(III), which often represents the most abundant electron acceptor in subsurface environments [9]. It is now recognized that not only vegetative cells, but in some instances spores, can catalyze U(VI) reduction [10[•],11]. Recently recognized U(VI)-reducing microorganisms include: *Pseudomonas* sp., *Pantoea* sp. and *Enterobacter* sp. recovered from the soil of a uranium mine [12]; several *Geobacter* species isolated from contaminated sites [9,13]; and the thermophile *Thermus scotoductus* [14[•]].

T. scotoductus contains a putative peptide ABC transporter, peptide-binding protein capable of U(VI) reduction [14[•]]. It was concluded that U(VI) reduction by this protein was fortuitous. Promiscuous reduction of U(VI) by proteins with other redox functions is probably a common theme in microbial U(VI) reduction, because, although microbial U(VI) can be driven by natural organic matter in sedimentary environments [15[•]], levels of U(VI) in uncontaminated waters are typically so low that it is unlikely that specific U(VI) respiratory pathways have evolved [3]. For example, reduction of U(VI) by *Geobacter* species appears to be quite non-specific, with a diversity of outer-surface

Figure 1



Conceptual illustration of the process of uranium bioremediation. **(a)** Indigenous microorganisms present in soils, sediments, and groundwater contaminated by nuclear energy and weapons production activities are stimulated through introduction of organic carbon compounds via injection wells. Select organisms may couple the oxidation of organic carbon (and H_2) to the reduction of aqueous uranium, as $U(VI)$, converting it from a soluble to an insoluble form, as $U(IV)$. **(b)** Reduced $U(IV)$ may be re-oxidized to $U(VI)$ following cessation of organic carbon injection accompanying subsequent delivery of oxidants, such as O_2 , NO_3^- , and Fe^{3+} ; the presence of diffusional barriers (e.g., biomass or low permeability sediments) or preferential reductants (e.g., FeS) can suppress re-oxidation and maintain stability of immobilized $U(IV)$.

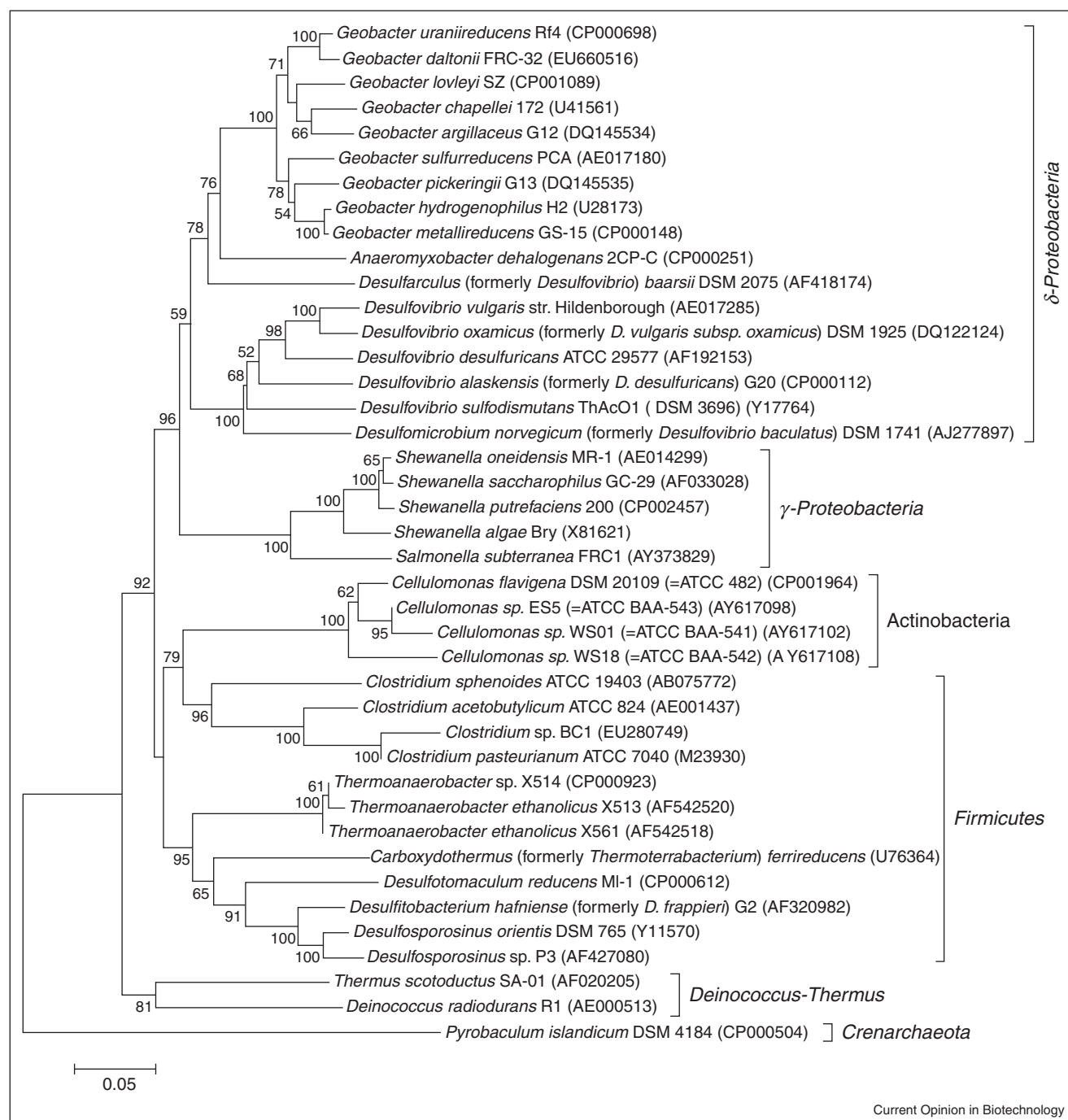
c -type cytochromes capable of transferring electrons to $U(VI)$ [9] in a manner similar to that reported for humic substances [16]. *Geobacter* species have the ability to reduce electron acceptors at substantial distance from the cell via pili with metallic-like conductivity [17,18] and an associated c -type cytochrome [19], but the suggestion that pili are the major site for $U(VI)$ reduction [20] is inconsistent with multiple lines of evidence that suggest that pili are not required for $U(VI)$ reduction [21].

It is difficult to definitively determine which microorganisms are responsible for $U(VI)$ reduction in most subsurface environments, as the availability of alternative electron acceptors capable of supporting anaerobic respiration, such as nitrate, $Fe(III)$, or sulfate [1[•],3], is typically much greater than that of $U(VI)$, even in contaminated environments. Thus, although it is common to ascribe the major role in $U(VI)$ reduction to the most dominant organisms in the groundwater during uranium bioremediation, in reality a minor component of the community may be sufficiently abundant to account for the observed $U(VI)$ reduction. Although many studies have focused on

planktonic populations and the $U(VI)$ in groundwater, microorganisms affixed to subsurface sediments can play an important role in $U(VI)$ reduction, as $U(VI)$ adsorbed onto sediments can also be reduced, converting it to forms less prone to desorption and release to groundwater [22[•],23] (Figure 3).

Many environmental conditions (e.g., type of organic substrates added, pH, salinity, etc.) influence which microorganisms predominate during *in situ* uranium bioremediation [24,25]. In some instances, molecular microbial community analysis has revealed rather complex microbial communities associated with bioremediation of uranium-contaminated groundwater [1[•]]. This was apparent in different treatment zones at the U.S. Department of Energy's (DOE's) Integrated Field Research Challenge (IFRC) site at Oak Ridge, Tennessee (USA). Groundwater at 'Area 3', which had low pH and high nitrate in addition to uranium contamination, was treated with an inner $U(VI)$ bioreduction loop, nested within an outer groundwater-conditioning loop. Geochip analyses revealed distinct microbial communities in the

Figure 2

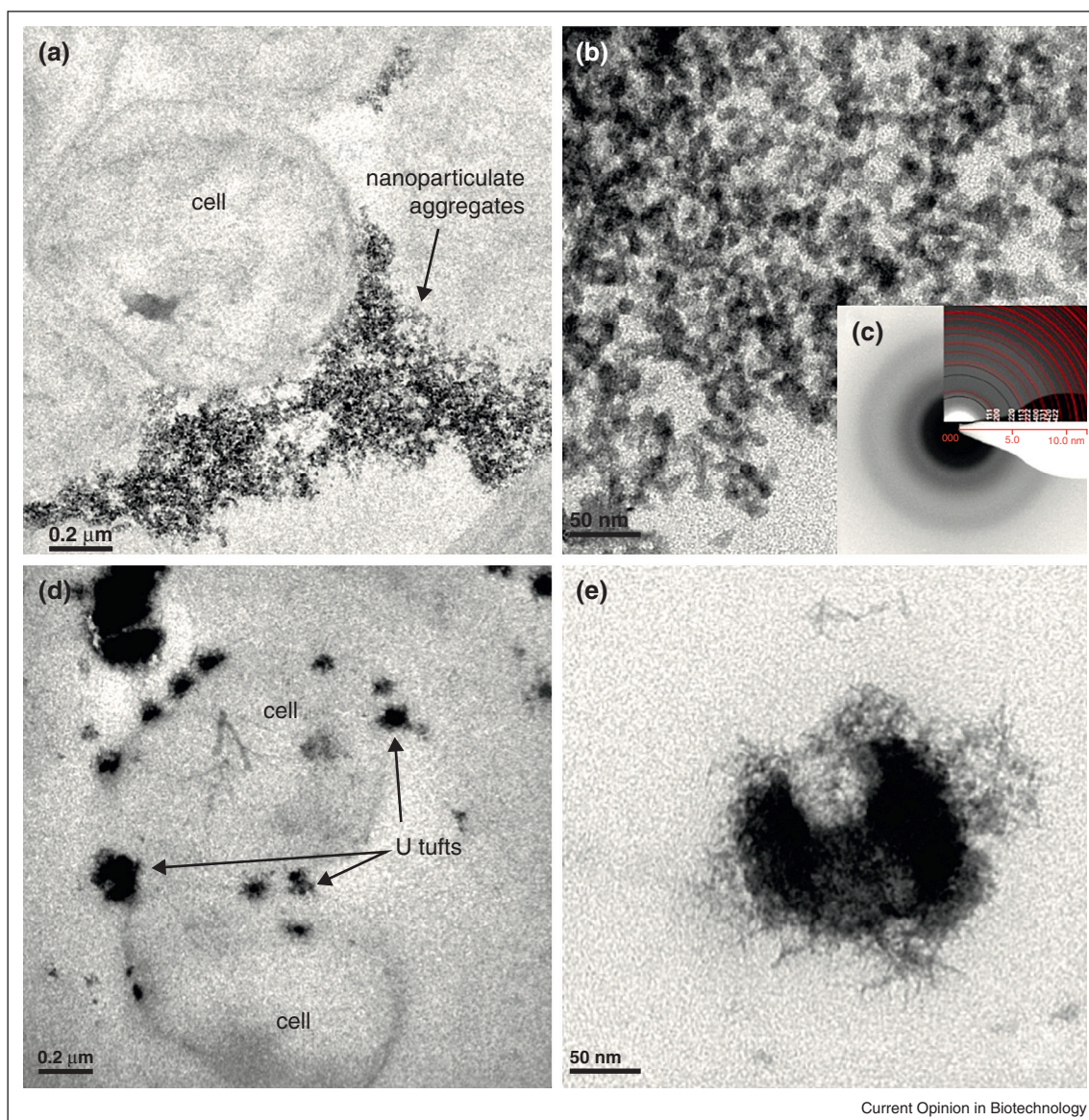


16S rRNA based phylogenetic tree of U(VI)-reducing prokaryotes. Adapted from Kostka and Green [1**].

two components of the treatment system, with metal-reducing bacteria, such as *Desulfovibrio*, *Geobacter*, *Anaeromyxobacter* and *Shewanella* species highly abundant in the inner *in situ* U(VI)-reduction zone, which received ethanol as an electron donor [26,27]. A massively parallel sequencing-indicator species approach identified *Desulfo-*

vibrio, *Anaeromyxobacter*, and *Desulfosporosinus* species as the predominant organisms [28]. Microorganisms in the families *Burkholderiaceae*, *Comamonadaceae*, *Oxalobacteraceae*, and *Rhodocyclaceae* predominated when microbial activity was stimulated with ethanol and methanol in sediment incubations [24].

Figure 3



Products of U(VI) bioreduction. Comparison of *Shewanella oneidensis* cells (a–c) and associated nanoparticulate uraninite (UO_2) to *S. oneidensis* cells exhibiting localized deposits of poorly ordered U(IV) coordination polymers (d,e); scale bars in (d,e) are 200 and 50 nm, respectively. U(IV) morphologies are distinct, with U(IV) coordination polymers staining tuft-like masses on the cells. Images are bright-field transmission electron micrographs; (c) is a selected area electron diffraction pattern interpreted as uraninite. Adapted with permission from Bernier-Latmani *et al.* [72••].

In studies at 'Area 2', which has circumneutral pH and lower nitrate levels, *Actinobacteria* were the most active under metal-reducing conditions in assimilating ^{13}C from the ^{13}C -ethanol added to promote microbial activity in laboratory sediment incubations [29]. Other laboratory studies with sediments from the Oak Ridge site demonstrated that the type of electron donor added and sediment origin greatly influenced which microorganisms predominated following addition of electron donors [24]. In a field study at this site, addition of emulsified

vegetable oil stimulated nitrate, Fe(III), U(VI), and sulfate reduction resulting in a succession of members of the *Comamonadaceae*, *Geobacteraceae*, and *Desulfobacterales* [30].

In contrast to the complexity of the microbial community at the Oak Ridge site, a much simpler microbial community has been noted during the most active phase of U(VI) reduction at the DOE's IFRC site at Rifle, CO, USA [2••]. This may in part be attributed to the circumneutral pH, negligible nitrate, and the fact that acetate, a

non-fermentable substrate, served as the electron donor to promote U(VI) reduction. Increasingly sophisticated molecular analyses performed using a field-portable microarray analysis system [31], Geochip [32] and PhyloChip technologies [33], environmental proteomics [34–36], and analysis of ^{13}C -acetate incorporation into lipids [37] and DNA [2^{••},22[•]] have confirmed the findings derived from earlier 16S rRNA gene sequence analyses that *Geobacter* species can account for up to 90% of the 16S rRNA sequences recovered from groundwater during active U(VI) reduction [3,9]. *Geobacter* species possess a number of physiological characteristics that may favor their growth in anaerobic subsurface environments [1^{••},9,38], including the recently recognized capability to grow as electron-donating partners in syntrophic relationships when Fe(III) is not available [39,40]. At the Rifle site, *Geobacter* species initially grow with Fe(III) as their primary electron acceptor; however, their ability to effectively reduce U(VI) [9] for sustained periods of time suggests that they may also be the primary U(VI) reducers when iron and sulfate reduction (and perhaps methanogenesis) are the predominant metabolic pathways, supplementing U(VI) reduction with reduction of elemental sulfur or syntrophic growth [2^{••}].

Systems analysis and genome-scale modeling of uranium bioremediation

The relative simplicity of the microbial community associated with uranium bioremediation at the Rifle site has made it feasible to analyze and model the *in situ* activity of important subsurface populations and develop tools that can now be potentially applied to more complex bioremediation sites. As recently reviewed [9], a series of studies quantifying transcript abundance of key genes in subsurface *Geobacter* populations elucidated important physiological parameters, such as rates of metabolism [2^{••}], growth rates [9], limitations for key nutrients, such as electron donor [41], phosphate [42], and fixed nitrogen [43], as well as other environmental stresses [43]. Environmental proteomic studies confirmed the abundance of *Geobacter* associated with peak rates of U(VI) reduction and provided information on relative metabolic activity [34,35,44]. The ability of *Geobacter* species to utilize electrodes as electron acceptors has made it feasible to also monitor their *in situ* rates of metabolism by emplacing electrodes in the subsurface [45].

Ideally, bioremediation strategies will be designed in the future with models that can accurately predict the outcome of bioremediation before field implementation, enabling optimization of bioremediation approaches. For example, it may be desirable to titrate in the minimum amount of electron donor necessary to promote effective U(VI) reduction, without producing excess biomass that may reduce permeability [2^{••}] or stimulate the growth of organisms, such as acetate-oxidizing sulfate reducers [46] that consume added electron donor but

do not contribute to U(VI) reduction. Models for describing U(VI) reduction in subsurface sediments are being developed with several different approaches [47^{••},48–51].

Development of truly predictive models will require the ability to forecast how key components of the microbial community respond to changing environmental conditions during bioremediation. One potential approach is Bottom-Up Genome Scale (BUGS) modeling in which genome-scale metabolic models of key organisms influencing the bioremediation process are coupled to geochemical and hydrological models [9]. An advantage of BUGS modeling is that it accounts for important changes in microbial physiology (e.g., growth yields) associated with the changing environmental conditions that microorganisms experience during uranium bioremediation; such features are not represented in traditional approaches for modeling microbial activity. BUGS modeling has been effective in predicting the course of uranium bioremediation in the relatively simple case of acetate-driven uranium bioremediation by coupling genome-scale models of *Geobacter* with reactive transport models [8,52–55,56[•],57].

BUGS modeling can describe more complex microbial communities with the successive addition of genome-scale models representing additional key populations. For example, it greatly benefits *in situ* uranium bioremediation that *Geobacter* species, which are capable of U(VI) reduction, outcompete acetate-oxidizing Fe(III)-reducing *Rhodoferrax* species (reportedly incapable of U(VI) reduction) when acetate is added to the subsurface. Following development of a *Rhodoferrax* genome-scale model [58], BUGS modeling predicted the ability of *Rhodoferrax* to effectively compete with *Geobacter* in the subsurface before acetate amendments and elucidated the physiological features that allow *Geobacter* to outcompete *Rhodoferrax* when high concentrations of acetate are made available [56[•]]. Modeling the interaction between *Geobacter* species and acetate-oxidizing sulfate reducers that compete for acetate, but do not reduce U(VI), improved the understanding of observed patterns of U(VI) reduction during acetate amendment [57] and suggested a strategy for enhancing uranium bioremediation through Fe(III) addition designed to maintain the long-term U(VI)-reducing activity of *Geobacter* species [59]. The ability of genome-scale metabolic modeling to predict mutations that change metabolic fluxes to adapt to new substrates may represent an additional predictive power for optimizing bioremediation strategies [60].

The BUGS modeling approach requires a substantial initial investment in isolating and characterizing microorganisms in order to obtain the physiological data required for modeling. In the future, this may be expedited by major advances in metagenomics that, as recently illustrated at the Rifle site, are providing

important basic information on uncultured organisms that should facilitate identification of the appropriate strategies for recovering these organisms in culture [61^{••}]. Ultimately, BUGS modeling can make the design of bioremediation strategies far more predictive and scientific, removing the substantial empiricism associated with the typical current practices for bioremediation of uranium and other contaminants.

Geochemical factors influencing uranium bioremediation

Also crucial for predicting the success of microbial U(VI) reduction as a remediation strategy is an understanding of the abiotic factors influencing the stability of reduced uranium species. At circumneutral pH, U(VI) in groundwater may primarily be complexed with inorganic carbon [2^{••}] in forms that are less susceptible to microbial U(VI) reduction than U(VI) in organic complexes [62,63], which are more readily reduced via enzymatic processes than with potential abiotic reductants. Although reduced iron [64,65] and sulfur [66,67] species can reduce U(VI) under laboratory conditions where inorganic constituents, such as Ca^{2+} and HCO_3^- , are included, incubation at environmentally relevant concentrations severely inhibits (or suppresses) reduction [68[•]]. This is consistent with field observations that U(VI) is only effectively removed when enzymatic processes are likely to be operative [2^{••}].

Although the most studied product of microbial U(VI) reduction is nanoparticulate (diameter <3-nm) uraninite [69–71], recent studies have reported poorly ordered coordination polymers of U(IV) coordinated to phosphoryl and/or carboxylate groups on biomass [72^{••}] or mineral phases, such as ningyosite, $\text{CaU}(\text{PO}_4)_2$ [71,72^{••},73^{••},74]. The chemical and physical forms of U(IV) and the geochemical environment influence the susceptibility of U(IV) to oxidation and mobilization via the formation of complexes [75–80]. While susceptibility of biogenic uraninite to oxidation is diminished by structural incorporation of Mn^{2+} [81] and reaction with Ca^{2+} [82], the most important stabilizing mechanisms (see Figure 1B) appear to be maintenance of low levels of dissolved oxygen and/or sequestration within low permeability sediments or enveloping biomass [76].

Concerns about potential re-mobilization of uranium precipitated as U(IV) illustrate the important limitation of microbial U(VI) reduction as a bioremediation strategy, which is that uranium remains in the subsurface. Other potential drawbacks are loss of aquifer permeability due to biomass and mineral accumulation [2^{••}], which can impede prolonged delivery of organic carbon, and the release of toxic contaminants, such as arsenic [83], adsorbed onto Fe(III) oxides.

An alternative that overcomes these limitations is to promote microbial U(VI) reduction by emplacing elec-

trodes that serve as the electron donor for the reduction of U(VI) [3,84]. Negatively poised electrodes provide a constant, steady source of electrons without promoting significant microbial Fe(III) or sulfate reduction and the deleterious water quality issues associated with those reactions. Furthermore, the U(IV) produced in electrode-driven microbial U(VI) reduction precipitates on the electrodes, facilitating removal of uranium from the subsurface. Although not predicated on reduction, alternative mechanisms for uranium bio-immobilization, such as biomineralization of insoluble U(VI)-phosphate species [85[•]] through microbial phosphatase stimulation, continue to attract interest for subsurface remediation.

Conclusions

Promoting microbial U(VI) reduction has proven to be an effective strategy for removing uranium from contaminated groundwater in the short term, thereby preventing its mobility. However, there is still significant uncertainty about the long-term sustainability of this approach and many options that might optimize the process have yet to be evaluated. An important outcome of the intensive investment in uranium bioremediation research in the last decade has been the development of approaches for assessing and modeling the activity of subsurface microbial communities that are expected to be applicable to the other forms of subsurface bioremediation, as well as to the study of diverse microbial processes in a wide range of soils and sediments.

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